1 ORIGIN, COMPOSITION, AND TRANSFORMATION OF

2 DISSOLVED ORGANIC MATTER IN TROPICAL PEATLANDS

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21 Abstract

22 Solid and Dissolved Organic Matter (DOM) compositions were investigated in a pristing 23 and a deforested tropical peat forest in Brunei Darussalam. A combination of elemental (%C, %N, C/N), isotopic (δ^{15} N, δ^{13} C, Δ^{14} C), molecular (lignin phenol biomarkers) and 24 25 optical (Specific UV Absorbance at 280 nm (SUVA₂₈₀), Fluorescence Index (FI)) 26 analyses were performed to characterize DOM in porewater and river water. The DOM 27 composition was compared to vegetation and peat to assess DOM origin and 28 transformations in the ecosystem. Significant relationships were observed between optical properties (SUVA₂₈₀, FI) and bulk (Δ^{14} C) and molecular (ratio of cynnamyl to 29 30 vanilly phenols, C:V) composition of DOM. 31 The radiocarbon content of DOM shows that it is modern at both sites for both surface and deep (to 4.5 m) samples (Δ^{14} C= 74.58±8.53% at the pristine site and Δ^{14} C= 32 33 87.84±4.5‰ at the deforested site), indicating transport of young DOM to deeper layers. 34 Stable carbon-13 content of vegetation, peat and DOM showed only slight fractionation 35 and ranged from -32.2‰ to -28.3‰. Nitrogen showed a greater shift between ecosystem pools. Nitrogen-15 content was higher in the solid peat ($\delta^{15}N = -0.9 \pm 0.4\%$ in the pristine 36 site, $\delta^{15}N = -1.2 \pm 1.4\%$ in the deforested site) than in DOM ($\delta^{15}N = -4.3 \pm 2.5\%$ in the 37 pristine site, $\delta^{15}N = -4.1\pm2.3\%$ in the deforested site). This difference in $\delta^{15}N$ is 38 39 significantly correlated with the C/N values of peat and DOM and attributable to higher 40 microbial degradation in the peat compared to DOM. Pore water DOM contains less lignin ($\lambda 8 = 1.16 \pm 0.35$ mg.100mg dw⁻¹ in the pristine site and $\lambda 8 = 1.51 \pm 0.51$ mg.100mg 41 dw⁻¹ in the deforested site) than does the solid peat ($\lambda 8 = 5.40 \pm 2.11 \text{ mg}.100 \text{mg} \text{ dw}^{-1}$ in the 42 43 pristine site and $\lambda 8 = 9.55 \pm 4.7 \text{ mg.} 100 \text{mg dw}^{-1}$ in the deforested site). All indicators of 44 lignin degradation (P:(V+S), 3.5Bd/V, (Ad/Al)V, (Ad/Al)S) are significantly higher in

DOM than in the solid peat. This shows that lignin can be processed rapidly in the porewater of upper layers of tropical peatlands.

Logging activity affected the composition of organic matter at the deforested site. Higher lignin content (λ8) was observed in the solid peat, and the composition of DOM differed greatly in the deforested site: Δ¹⁴C, FI, 3.5Bd/V, (Ad/Al)V are significantly different from the pristine site. Thus it appears that the composition of DOM is more sensitive than the solid phase to the effects of land use change on organic matter dynamic in tropical

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peatlands.

- 54 **Key-Words:** Dissolved organic matter, Tropical peatland, Stable isotopes, Radiocarbon
- dating, Lignin biomarkers, Absorbance, Fluorescence

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1. Introduction

58 Dissolved organic matter (DOM) constitutes a small but important fraction of soil carbon 59 stocks because it is both reactive and mobile. Within porewater, DOM performs several 60 ecological functions: provision of carbon and energy for microorganisms, production of 61 CO₂ and CH₄ that may be released to the atmosphere, and control of porewater pH and speciation of diverse nutrients (N and P) and metallic elements (Neff et Asner 2001; 62 Limpens et al. 2008). In peatlands, dissolved organic matter (DOM) is also a key 63 64 component of the carbon cycle. Transfer of young DOM from surface to deep layers can 65 induce methanogenesis and/or microbial decomposition in temperate peatlands, 66 producing CO₂ and CH₄ deep in the peat (Aravena et al. 1993; Chanton et al. 1995; 67 Chanton et al. 2008). Recently, fluvial export of dissolved organic carbon (DOC) has 68 been recognized as a significant component of the carbon budget of both temperate and 69 tropical peatlands (Billett et al., 2010, Moore et al., 2013). DOM exported to streams may 70 decompose there, becoming a source of CO₂ to the atmosphere (Cole et al. 2007). Tropical peatlands export dissolved organic carbon at a higher rate (above 80 g.m⁻².y⁻¹ 71 72 (Moore et al., 2011; Gandois et al., 2013)) than do temperate and boreal peatlands (10 to 25 g m⁻² y⁻¹, (Fraser et al., 2001)) so that, although they constitute only 11% of peatlands 73 74 by area (Page et al., 2011), tropical peatlands may nonetheless release a significant 75 proportion of all peatland dissolved organic carbon exports. Despite the recent evidence of its crucial role, the composition and fate of peatland DOM 76 77 remain uncertain, although it is expected to be very diverse because of the likely diversity 78 in its origin and transformation (vegetation, peat, microbial activity). Elucidating DOM 79 composition and fate is crucial to understanding its role in both peatland carbon dynamics 80 and the global carbon cycle. 81 Most tropical peatlands (77%) are located in South-East Asia (Page et al., 2011). Before 82 development, these peatlands were mostly forested and hosted significant biodiversity 83 (Posa et al., 2011). Tropical peatlands have been rapidly degraded due to drainage, 84 deforestation and conversion to agriculture since the late 1990's (Miettinen et al., 2011). 85 This abrupt change in land use has induced massive release of carbon from peatlands to 86 the atmosphere because of enhanced decomposition of organic matter (Couwenberg et al., 87 2009; Hergoualc'h et Verchot, 2011). The modification of organic matter from peatland 88 degradation has also affected DOM composition and export. Evidence of 89 anthropogenically induced changes in the quantity, composition and age of exported

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DOM in temperate peatlands has been reported (Kalbitz 2001; Freeman et al. 2001; Pastor et al. 2003; Evans et al., 2007; Billett et al. 2010). Recent studies suggest that similar changes have occurred in tropical areas (Gandois et al., 2013; Moore et al., 2013). Investigation of the composition of DOM is necessary to understand its role in the carbon cycle and its response to anthropogenic pressure in tropical peatlands. In the work presented here, complementary tools (isotopic, molecular, optical) were applied to different ecosystem compartments (DOM, solid organic matter, vegetation) to investigate tropical peatland organic matter dynamics. Isotopic shifts (δ^{13} C and δ^{15} N) between the different pools of an ecosystem indicate organic matter transformation (DeNiro et Hastorf 1985; ŠantRůČková et al., 2000). The fate of lignin is of primary importance for understanding tropical peat accumulation because tropical peat originates from trees (Durig et al. 1988; Calvert et al., 1991). We assessed lignin source and degradation in vegetation, peat and DOM samples by analysis of CuO oxidation products (Thevenot et al., 2010). Upon oxidation (alkaline CuO), the lignin macropolymer is hydrolyzed to small structural units belonging to three classes of methoxylated phenols: vanillyl (V), ubiquitous in all terrestrial plants, syringyl (S), specific of angiosperms and cinnamyl (C), produced in non-woody tissues. The yields and ration of these phenols can be used to identify lignin origin (Hedges et Mann 1979). Each class of lignin oxidation products in turn comprises an acid, an aldehyde and a ketone (for V and S) or only acids (for C). The ratios of compounds reflect lignin transformation in terrestrial environments arising from microbial activity (Goñi et al. 1993; Dittmar et Kattner 2003), photodegradation (Benner and Kaiser, 2011) or sorption to soil phases (Hernes et al., 2007). This is the first time this tool has been used to study

both solid and dissolved organic matter in tropical peatlands. Investigations of optical properties of DOM provide information on DOM composition, mainly aromatic molecules (Weishaar et al. 2003), origin and transformation in ecosystems (McKnight et al. 2001; Yamashita et al. 2010). We have combined these indicators in the context of a pristine and a deforested tropical peat dome of Brunei to investigate: (1) DOM and lignin origin and transformation in tropical peatlands, (2) Influence of deforestation on organic matter and DOM in tropical peatlands.

2. Material and methods

2.1. Study sites

The study sites are located in the Ulu Mendaram Conservation Area, Belait District, Brunei Darussalam, on the Island of Borneo (Figure 1). Two peat domes were investigated. One is a pristine peat dome (Mendaram dome) and the second (Damit dome) is a former timber concession, where selective logging, but no drainage, occurred from 1980 to 2010. At the pristine site, the main vegetation cover includes *Shorea albida* (dipterocarp species listed in the IUCN (International Union for Conservation of Nature and Natural Resources) red list of threatened species) and *Pandanus* sp. (pandan) as described in the early 1980 (Anderson, 1983). In the deforested site, most of the *Shorea albida* has been harvested, with the exception of a few individuals. Recovery species include *Pandanus* sp. as well as *Nephrolepsis biserrata* (ferns) (Kobayashi, 1999).

2.2. Sample collection

133 -Peat samples:

Samples were collected at 3 locations along transects from the edge (river) to the center of each dome. At each location, 3 replicates were sampled in a 10 m x 10 m square. Surface peat was collected to 15 cm depth after litter removal. Samples were placed in plastic bags, transported to the laboratory, and dried at 60 °C. Roots were removed and the remaining dry material was ground in an agate mortar.

- Cores

Three cores were collected at each site for radiocarbon dating of the bulk peat using a 5 cm diameter Russian peat corer (Edelman auger, Eijkelkamp, Netherlands). The cumulative length of each retrieved core varied from 3 m at the edge of the domes to 6 m at the center. For dating purposes, three subsamples were collected on each core between 65-69 and 126-130 cm depth and one at the bottom of the peat, before the colour and texture transition with the clay, the depth of which varied with the location of the core on the peat dome. Deep samples were taken at 245-248, 507-510,531-535 cm depth in the pristine site and 187-190, 383-386, 424-427, 451-454 cm in the logged site. Samples were dried (60°C) and ground in an agate mortar.

- Water samples

Water was sampled in the middle of the river draining the two sites (Mendaram River) and in the river draining the north side of the Damit dome (Damit River). Samples were taken at a 20 cm depth using cleaned polypropylene bottles. Porewater was sampled along a transect from the edge to the center of the dome. The 30 cm and 180 cm depth samples were extracted using inox pushpoint samplers (MHE products, USA), connected to Tygon tubing (St Gabon, France) and a hand pump. After flushing the system, a sample was collected using a polypropylene syringe. At each location on the domes,

water was also collected from pools at the peat surface. Sampling equipment was cleaned with acid (HCl, 1N) and ultrapure water. Glass bottles, filtration apparatus and filters were baked (550°C) for 43 hours to avoid any carbon contaminations. Samples were filtered within 24 h. An aliquot was stored in glass bottles for total organic carbon (TOC) and optical properties analysis. One liter was freeze-dried to collect dissolved material for isotopic and molecular analysis. Deeper pore water samples were collected at one location in the pristine site using the same sampling protocol as the shallow samples. The samples were not freeze dried and direct analysis of Δ^{14} C was performed on the filtered sample (acidified with ultrapure HCl). Blank showed no carbon contamination.

-Vegetation samples:

Vegetation was sampled at both sites to investigate the main sources of organic matter in the peat. *Pandanus* leaves (3 replicates at both sites) and *Shorea albida* leaves, bark and wood (2 replicates) were sampled. In the deforested site, *Nephrolepis biserrata* (2 replicates), abundant after deforestation, was also sampled.

2.3. Sample analysis

- Porewater organic carbon content
- Non-purgeable organic carbon (NPOC, referred to hereafter as DOC) was analyzed on
- 174 filtered samples after acidification to pH 2 with a TOC-V CSH analyzer (Shimadzu,
- 175 Japan). Certified material (soft river water, Miramichi-02, Environment Canada) was
- included in the analytical loop and recovery was >95% of the certified value.
 - Optical properties of DOM

178 The UV absorption spectra from 240 to 700 nm of porewater were measured with a 179 spectrophotometer (UV-1601 PC, Shimadzu, Japan) in a 1 cm quartz cell. The baseline was determined with ultra pure water. The Specific UV Absorbance at 280 nm (SUVA₂₈₀, 180 $1.\text{mg}^{-1} \text{ m}^{-1}$) was calculated as follows: $SUVA = \frac{A_{280}}{h.[DOC]} (1.\text{mgC}^{-1}.\text{m}^{-1})$, where A_{280} is the 181 182 sample absorbance at 280 nm (non-dimensional), b is the optical path length (m) and 183 DOC is in mg.1⁻¹. The absorbance at 280 nm was chosen instead of the usual 254 nm 184 (Weishaar et al., 2003) since the high DOC concentration induced saturation of the 185 spectrophotometer at 254 nm. Fluorescence measurements were performed on porewater samples with a LS 45 Perkin 186 Elmer (Norwalk, CT, USA) fluorescence spectrometer. Emission slits were set at 5 nm 187 with a scan speed of 500 nm min⁻¹. Emissions spectra were recorded from 350 to 700 nm 188 189 for 370 nm excitation. The intensity of the fluorescence spectra was corrected for inner 190 filtering effect using the absorbance spectra following the procedure described in 191 McKnight et al. (2001). Fluorescence Index was calculated as defined in Maie et al. (2006), by the ratio of the fluorescence intensity at 470nm to intensity at 520 nm for a 192 193 370 excitation. 194 -Isotopic analysis Total C and N content, as well as stable isotope analyses (δ^{15} N and δ^{13} C) of DOM and 195 196 peat samples were performed using a mass spectrometer (Isoprime 100) coupled with a Vario El Micro elemental analyzer. $\delta^{15}N$ and $\delta^{13}C$ were calculated using (R_{sample}/R_{standard}-197 1) x 1000, where R is the ratio of ¹³C/¹²C, or ¹⁵N/¹⁴N. Standard was Pee Dee Belemnite 198

- for carbon and atmospheric N₂ for nitrogen. For each analytical sequence, standard and replicates were included.
- 201 Radiocarbon analysis of DOM and bulk peat was performed at NOSAMS (Woods Hole,
- 202 USA) AMS facility following standard methods.
- 203 Molecular biomarkers analysis

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Molecular biomarker analysis of DOM and peat samples was performed according to the copper oxidation (CuO) method initially developed by Hedges and Ertel (1982) and modified by Goñi and Montgomery (2000). 3-5 mg of sample (vegetation, peat or freezedried DOM) were oxidized with CuO in a 2 mol L⁻¹ NaOH solution under exclusion of O₂ in an oven at 150°C for 180 min in 3.2 mL stainless steel mini-bombs (Prime Focus® Inc, Seattle). Once vials were opened, ethyl vanillin and cinnamic acid were added to the solutions as an internal standard. The aqueous solutions were acidified to pH 1 with concentrated HCl and extracted with ethyl acetate. After evaporation in a centrivap cold trap, reaction products were dissolved in pyridine and derivatized with bis(trimethylsilyl) - trifluoroacetamide-trimethylchlorosilane (BSTFA-TMCS) to form trimethylsilyl derivatives. Extraction products were then analyzed on a Gas Chromatograph-Mass Spectrometer system (Model 3800-Saturn 2000, Varian) fitted with a fused capillary column (Varian FactorFour VF-1ms 60 m, 0.32 mm). Each compound was identified according to its retention time and by comparing it to commercially available standards. Standards were analyzed every 4 samples, and the standard deviation of replicate analysis was below 15%. We analyzed 12 CuO derivatives used in recent literature, three lignin derived phenols families: Vanillyl phenols (V: VI, vanillin; Vn, acetovanillone; Vd, vanillic acid), cinnamyl phenols (C: p-Cd, p-coumaric acid; Fd, feluric acid) and syringyl

phenols (S: Sl, syringaldehyde; Sn, acetosyringone; Sd, syringic acid). These three families are used to determine the source of organic matter along with relative abundance (λ8: sum of ΣV, ΣC and ΣS). P-hydroxyphenols (Pl, p-hydroxybenzaldehyde; Pn, p-hydroxyacetophenone; Pd, p-hydroxybenzoic acid) and a benzene carboxylic acid (3,5-dihydroxybenzoic acid) were also combined with V and S to describe the TOM state of degradation and humification (Houel et al 2006, Dickens et al 2007).

3. Results

Results are summarized in Table 1 for bulk analysis of organic matter and in Table 2 for organic matter biomarkers. Henceforth DOM and peat samples from pristine and deforested domes have been pooled. DOM elemental, isotopic and molecular compositions did not show significant differences (Kruskal Wallis, p<0.001) between sampling sites within the same dome and within vertical profiles (flat profiles of DOM characteristics between surface, 30 cm and 180 cm samples).

3.1. Bulk elemental, stable and radiocarbon isotopic composition of peat,

DOM and vegetation

237 • Elemental content

DOC concentrations were high in the peat porewater: $66.1\pm8.9 \text{ mg.l}^{-1}$ in the pristine site and $74.9\pm9.1 \text{ mg.l}^{-1}$ in the deforested site (Table 1). Lower DOC concentrations were measured in the river water: $37.4\pm0.1 \text{ mg.l}^{-1}$ in the pristine river separating the two sites, which receives input primarily from peatlands, and 26.7 in the deforested river draining the north side of the deforested site which has headwaters beyond the peat lands. The carbon concentration in samples was lower in DOM samples (C = 40 to 47 %) than in

- 244 vegetation (C= 46 to 52%) and peat samples (C = 48 to 56 %) (Table 1). Nitrogen content 245 ranged from 0.6 to 1.1% in DOM samples, from 0.8 to 1.4% in vegetation (except in 246 ferns (N= 2-3%)) and from 1.8 to 3% in peat samples. Nitrogen was enriched in peat 247 samples (C/N = 30.6 ± 5.7 compared to DOM (C/N= 53.2 ± 10) and vegetation samples, 248 which exhibited low N content and high C/N ($C/N=55.0\pm12$), except for ferns (Table 1). 249 Stable isotopes 250 The most abundant plant species at the pristine site (Shorea albida and Pandanus sp.) showed similar carbon isotopic ratios ($\delta^{13}C = -29.6 \pm 0.04\%$, Figure 2a), typical of C₃ 251 252 vegetation. The ferns (Nephrolepis biserrata) that recolonize the deforested site were relatively ¹³C depleted (δ^{13} C =-30.8±0.2‰). The isotope ratio of peat samples from the 253 pristine site was on average slightly 13 C depleted (δ^{13} C = -30.5±0.8‰) compared to the 254 deforested site (δ^{13} C = -29.5±0.6‰). In porewater, δ^{13} C-DOM values were homogeneous 255 with values observed in the peat and vegetation samples with an average of $\delta^{13}C = -$ 256
- 29.9 \pm 0.24‰ in the pristine site and δ^{13} C = -30.2 \pm 0.2‰ in the deforested site. The δ^{13} C258 DOM values of river water was similar in the Mendaram river (δ^{13} C = -29.6 \pm 0.05 ‰) and
- 259 in the Damit river (δ^{13} C = -29.9‰).
- 260 A wide range of $\delta^{15}N$ values was measured in the vegetation (from $\delta^{15}N$ = -3‰ to -8‰)
- and DOM ($\delta^{15}N$ = -0.8 ‰ to $\delta^{15}N$ = -8.2‰). The $\delta^{15}N$ -DOM ($\delta^{15}N$ = -4.18±2.5‰ in the
- pristine site, $\delta^{15}N$ = -4.13±2.3‰ in the deforested site) was lower than in the peat ($\delta^{15}N$ =
- 263 -0.9 \pm 0.4‰ in the pristine site, $\delta^{15}N = -1.2\pm1.4$ ‰ in the deforested site). A significant
- 264 (p<0.05) increase in $\delta^{15}N$ was observed with C/N decrease when pooling the peat and
- 265 DOM samples (Figure 2b).
- o Radiocarbon

Radiocarbon (Δ^{14} C-DOM) values in porewater and river were all modern (Table 1). All 267 the values were greater than the atmospheric value at the time of sampling ($\Delta^{14}C = 50\%$ 268 in 2008, (Graven et al., 2012)). Averages of Δ^{14} C-DOM = 74.58±8.53‰ and Δ^{14} C-DOM 269 $= 87.84 \pm 4.5\%$ were measured in the pristine (first 3 meters) and in the deforested site. In 270 the pristine site, where additional samples were taken, a slight increase of Δ^{14} C-DOM is 271 272 noticed after 3 meters, but DOM remains modern down to 4.5 meters. In the river draining the pristine site, values of Δ^{14} C-DOM = 62.7 and 73.84% were measured. 273 274 Radiocarbon content of bulk peat organic carbon showed a sharp decrease with depth for 275 all cores. Average ages of 254 and 620 yr were measured at 65 and 128 cm depth, respectively for both sites. The deepest peat showed the greatest variability in age: at the 276 bottom of the cores, the measured age of the bulk material ranged from 2230 to 2920 yr. 277 Higher variability was also observed for the shorter (2 to 3 m) cores sampled on the river 278 279 edges (Figure 3).

3.2. Optical properties of DOM

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Optical properties of DOM in porewater and river water of the two sites are described in Gandois et al. (2013). Measured SUVA₂₈₀ values range from 3.4 to 6, with significantly (KW, p<0.05) greater values in the deforested site (5.2 ± 0.6 L.mgC⁻¹ m⁻¹) compared to the pristine site (4.3 ± 0.6 L mgC⁻¹ m⁻¹), indicating that DOM is more aromatic in the deforested site. Fluorescence index ranged from 1.33 to 1.62, reflecting the predominance of terrestrially-derived DOM (McKnight et al., 2001). Fluorescence index was significantly (KW, p<0.05) higher in the pristine site (FI= 1.41 ± 0.05) than in the deforested site (1.36 ± 0.04).

3.3. Lignin phenols content

o Total yield

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- All the lignin phenols measured in the samples are presented in Table 2. The total yield
- 292 ($\lambda 8$) is the sum of $\Sigma C + \Sigma S + \Sigma V$ (detail $\Sigma C + \Sigma S + \Sigma V$). Higher yields were measured in
- vegetation and peat samples than in DOM. A significant (Kruskall Wallis test, $\alpha < 0.05$)
- 294 increase in total lignin yield was observed in peat in the deforested site ($\lambda 8 = 9.55 \pm 4.7$
- 295 mg.100mg OC⁻¹) compared to the pristine site ($\lambda 8 = 5.40 \pm 2.11$ mg.100mg OC⁻¹). This
- increase was also observed in the dissolved phase ($\lambda 8 = 1.16 \pm 0.35 \text{ mg.} 100 \text{mg OC}^{-1}$) in the
- 297 pristine site and $\lambda 8 = 1.51 \pm 0.51$ mg. 10mg dw⁻¹ in the deforested site).
- 298 o S:V vs C:V
- 299 The cinnamyl:vanillyl (C:V) and syringyl:vanillyl phenol (S:V) ratios showed very
- 300 distinct values for different vegetation types (Figure 4). Pandanus leaves showed high
- 301 S:V ratio (>1) but low C:V ratio (<0.03). On the contrary, fern samples showed very low
- 302 S:V ratio (<0.2) but higher C:V ratio in leaves (>0.2). The Shorea samples showed
- intermediate S:V ratio (0.6 to 1.4) and higher C:V ratios in leaves than in wood samples
- 304 (0.34 vs. 0.004). The S:V (from 0.3 to 1.2) and C:V (from 0.3 to 1.2) ratio measured in
- 305 the peat and DOM at both sites were close to measurements in *Shorea albida* wood. The
- 306 S:V and C:V ratios were significantly correlated in DOM samples (S:V=-0.065C.V-
- 0.024, n=13, r²=0.67, p<0.001). No relationship was observed in the peat samples.
- \circ (Ad/Al)V vs (Ad/Al)S
- 309 Vanillic acid to vanillin ((Ad/Al)V) and syringic acid to syringaldehyde ((Ad/Al)S)
- values increased from vegetation to peat and DOM (Figure 6). Measured ratio in *Shorea*
- 311 and Pandanus samples showed similar values ((Ad/Al)S= 0.40±0.01 and

(Ad/Al)V=0.40±0.10). Ratios in peat samples ranged from 0.4 to 1 for (Ad/Al)V and 0.4 to 0.8 for (Ad/Al)S . The values were linearly correlated in peat samples ((Ad/Al)V=1.3*(Ad/Al)S – 0.03, n=18, r^2 =0.66, p<0.001)). The ratios (Ad/Al)S and (Ad/Al)V were not correlated in DOM as they were in peat samples and their values were significantly higher than in peat samples. In DOM samples, syringic acid to syringaldehyde ratios ((Ad/Al)S) were similar in the pristine (1.16±0.08) and the deforested site (1.18±0.05). In contrast, vanillic acid to vanillin ((Ad/Al)V) ratio was significantly (α <0.001) higher in the pristine (1.42±0.12) compared to the deforested site (1.05±0.15). In the river, ratios were close to those measured in porewater of both sites

321 o P:(V+S)

The ratio P:(V+S) was low in vegetation samples, except for the fern leaf samples (P:(V+S)= 1.4), related to the very low V content measured in this sample (Figure 5). This ratio greatly increased from peat samples $(0.29\pm0.05 \text{ in the pristine site}, 0.25\pm0.12 \text{ in the deforested site})$ to DOM samples $(1.09\pm0.22 \text{ in the pristine site}, 0.86\pm0.10 \text{ in the deforested site}, 0.8\pm0.02 \text{ in the river})$.

327 ○ 3,5Bd:V

The 3,5-dihydroxybenzoic (3.5Bd) was normalized to V content and plotted in relation to P:(V+S) ratio. The 3,5Bd:V ratio value was very low in the vegetation samples (<0.15), except in the *Pandanus* leaves (average value of 0.6), (Figure 5). Values in the peat were also very low (0.12 \pm 0.03 in the pristine site, 0.08 \pm 0.04 in the deforested site) and dramatically higher in DOM, similar to the pattern for P:(V+S). In DOM, a significantly higher (Kruskall-Wallis, p<0.005) value was observed in the pristine site (3,5Bd:V =

1.58 \pm 0.25) than in the deforested site (3,5Bd:V = 0.92 \pm 0.18). In the river, an intermediate value of 1.31 \pm 0.12 was measured.

4. Discussion

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4.1. Dissolved organic matter origin in tropical peatlands

The radiocarbon (Δ^{14} C) measured in porewater DOM is uniformly modern. Thus, the DOM must contain a fraction of carbon incorporated during the period of enhanced atmospheric ¹⁴C from nuclear bomb testing and can contain only a minor component of older carbon. That we find little influx of old organic carbon from upwelling groundwater is not surprising because the entire peat dome experiences net recharge, with downward flow from the surface. However, it is striking that DOM from our deepest 4.5m pore water samples is modern because the peat at this depth is hundreds of years old (Figure 3). Thus, radiocarbon-young dissolved organic carbon is found in a matrix of older organic peat, showing that DOM at our sites originates from above, from vegetation, litter or peat leaching; not from the peat at the same depth. Export of modern DOC from peatlands has been reported in boreal (Raymond et al., 2007), temperate (Evans et al., 2007, Billett et al., 2012, Garnett et al., 2012) and tropical areas (Moore et al., 2013). Other studies have also shown that most DOM exported from undisturbed peatlands originates from the young, upper layers of peat, but disturbance by intense drainage can induce the release of very old carbon, as observed in temperate (Kalbitz, 2001) and tropical (Moore et al., 2013) peatlands. This has important implications for the deep carbon dynamics of tropical peat. In temperate peatlands, transport of young DOM to deep layers has been shown to induce microbial activity (Chasar et al., 2000) and

356 methanogenesis (Chanton et al. 1995; Aravena et al. 1993; Chanton et al. 2008) at depth. 357 The consequences for tropical peatlands have yet to be evaluated. So far, measured 358 methane emissions are low compared to the temperate and boreal peatlands, but major 359 uncertainties remain as to the reasons (Couwenberg et al., 2009). 360 Stable carbon isotopic ratios measured in the vegetation (Shorea, Pandanus) samples 361 were typical of contemporary C₃ vegetation. Stable nitrogen isotopic ratios varied much 362 more in vegetation samples. This variability could be related to differing nitrogen fixation 363 strategies (including atmospheric fixation, or use of organic nitrogen) developed by plants 364 in peatlands, which are nitrogen-poor environments (Asada et al., 2005). In contrast to the deforested site, slightly lower δ^{13} C values were observed at the pristine site in peat 365 366 samples compared to vegetation. In soils and peatlands, microbial decomposition of organic matter usually induces δ^{13} C enrichment (Natelhoffer and Fry, 1988; Powers and 367 368 Schlesinger 2002; Boström et al., 2007). Lower observed δ^{13} C values in solid peat of the 369 pristine site could also arise from selective enrichment of lignin in the peat because lignin 370 is ¹³C-depleted relative to bulk organic matter (Benner et al. 1987; Gleixner et al. 1993);. 371 The higher $\delta^{15}N$ value in solid peat samples at both sites compared to vegetation samples 372 could relate to isotopic enrichment of organic matter following microbial processing of 373 litter as suggested by numerous observations in the organic layer of soils (Natelhoffer and 374 Fry, 1988; Kramer et al., 2003) and sphagnum peatlands (Martínez Cortizas et al., 2007). 375 At both sites, the δ^{15} N value for DOM was lower than for peat, and similar to vegetation (Figure 2a and 2b). The relationship between $\delta^{15}N$ and C/N values in different pools 376 showed that increase of $\delta^{15}N$ in peat and DOM was related to the decrease of the C/N 377 378 ratio, meaning that loss of carbon due to microbial transformation was related to a

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preferential loss of ¹⁴N (Malmer et Holm 1984; Kokfelt et al., 2009). The contribution of throughfall and litter on DOM composition in pore water has already been highlighted in mineral forests (Fillion et al., 1998; Hou et al., 2005) but not yet in peatland forests, where the most obvious source of DOM is peat itself.

4.2. Lignin transformation in tropical peatlands

Tropical peats, mainly derived from wood (Anderson 1983) are mostly composed of lignin and its biodegradation products (Durig et al., 1988; Calvert et al., 1991; Williams et al. 1998). Our analysis of lignin oxidation products in vegetation, solid peat OC, as well as porewater and river DOM, in a tropical peatland provides the first assessment of lignin transformation in these ecosystems. The lignin yields $\lambda 8$ and $\lambda 11$ ($\lambda 8+\Sigma P$) quantify the phenols analyzed following CuO oxidation but underestimate total lignin content (Williams et al. 1998). The yields measured are in the highest range of values measured in soils (reviewed by Thevenot et al., 2010), highlighting the high concentration of lignin in tropical peatlands. Cinnamyl:vanillyl (C:V) and syringyl:vanillyl phenol (S:V) ratios are often used as indicators of vegetative origin of the lignin fraction, because ratios are distinct for several large groups of plants (Hedges and Mann 1979). C:V and S:V ratios measured in the peat samples were similar to those measured by Tarek et al., (2004) in a core collected from a peat swamp in Java (low values of C:V < 0.2 and S:V < 1.5). The variability of each of these ratios in our surface samples was similar to that in the whole core. Comparison of lignin phenol composition between peat and Shorea albida wood and leaves suggest that peat is mostly composed of wood, but can contain some pandan. In sphagnum-dominated

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temperate peatlands, (Williams et al. 1998) and (Williams and Yavitt, 2003) report similar S:V values but higher C:V values (up to 0.38±0.4), in accordance with the higher content in non-woody lignin. Numerous lignin phenol ratios can be used to assess the extent of lignin transformation. Microbial lignin transformation mainly consists of demethylation of the aromatic phenol. Demethylation therefore reduces the yield of methoxylated phenols (vanilly) and syringyl phenols) but does not affect the yields of phenols without methoxyl groups (p-hydroxyl phenols). Therefore, the ratio P:(V+S) can be an indicator of lignin transformation by microorganisms (Dittmar et Kattner, 2003). Photodegradation of DOM can also lead to an increase of the P:(V+S) ratio because higher decay rates have been observed for V and S phenols compared to P phenols (Benner and Kaiser, 2011). The origins of benzene carboxylic acids like 3,5-dihydroxybenzoic acid (3,5Bd) are diverse in terrestrial environments. The main source in soils is lignin phenol transformation, although they may also originate from tannins and flavonoids (Dickens et al., 2007). The ratio 3,5Bd/V is used as an indicator of lignin transformation in various environments (Houel et al., 2006, Amon et al., 2012). The acid to aldehyde ratios of vanillin and syringyl phenols are also transformation indicators because they increase as a result of propyl side chain oxidation (Ertel et Hedges, 1984), although these ratio can increase from litter to DOM during solubilization. The values of the indicators of lignin transformation ((P:(V+S), 3,5Bd:V) did not vary greatly between vegetation and solid peat samples. The (Ad/Al)V and (Ad/Al)S ratios were only slightly higher in peat than in vegetation. This indicated only a limited transformation of lignin from litter to peat or leaching from litter. Similar conservation of lignin structure has been observed for the whole peat column by Tarek et

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al. (2004). In DOM, a substantial decrease in lignin yield (λ8) was observed, as well as a consistent increase in all indicators of lignin transformation (P:(V+S), 3,5Bd:V, (Ad/Al)V, (Ad/Al)S)), (Figure 5 and 6), compared to the solid peat and vegetation at the both sites. They all indicate fungal (mostly white rot fungi) and bacterial degradation (Hedges et al 1988, Goñi et al. 1993, Opsahl et Benner 1995) as well as photodegradation (Benner and Kaiser, 2011) of lignin of DOM in the surface layers. Together, these indicators of more advanced decomposition of DOM demonstrate that transformation of lignin can occur much faster in porewater than in solid peat. Whereas in the solid peat lignin is preserved for centuries and millennia (Tareg et al., 2004), it is transformed in the dissolved phase within years. The transformation might happen mostly at the very surface of the peat (both photo-oxidation and microbial degradation), because the O₂ content decreases sharply in the porewater below 10 cm (Gandois et al., 2013). Lignin, usually considered as one of the most refractory compounds of organic matter can be decomposed quickly in the vadose zone. The transformed lignin is then likely advected downwards in the porewater. The conditions for accumulation of tropical peatland are not only the high content of lignin (Yule et Gomez 2009), but the environmental factors, including anoxia, and to a lesser extent darkness and low pH.

4.3. Link between optical, bulk isotopic and molecular properties of DOM

Optical properties of DOM are useful for assessing its composition. Easy to perform and non-destructive, these analyses have been performed at ecosystem (Yamashita et al., 2010) and country (Butman et al. 2012) scales. Here we have combined the optical approach with more expensive and labor-intensive isotopic and molecular analyses to

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assess their potential for providing detailed information on DOM (Tfaily et al. 2013). Here, in this homogenous ecosystem, we observed that optical properties of DOM were related to molecular properties, which greatly expands the potential usefulness of fluorescence measurement of DOM. Although lignin analysis and optical properties refer to different fractions of DOM, lignin composition can be reconstructed based on fluorescence spectra (Hernes et al. 2009). We observed a significant correlation between the C:V and S:V ratio and the fluorescence index (Figure 7a). The correlation indicated that lower C:V and S:V ratios in the lignin fraction of DOM were related to DOM containing more carbon of terrestrial origin, as indicated by lower FI (McKnight et al., 2001; Yamashita et al., 2010). The radiocarbon content of DOM in pore and river water was also related to its optical properties in our samples: Δ^{14} C-DOM showed a significant negative relationship with FI (Figure 7b) and a significant (r²=0.37, p<0.01, n=11) positive relationship with SUVA₂₈₀. The higher values of Δ^{14} C, indicating lower residence time of DOM were related to lower fluorescence index and higher SUVA280 values, indicating less microbial processing of DOM and higher aromaticity. Correlation between DOM radiocarbon content and water optical properties was also reported by Sanderman et al., (2009) in a coastal watershed of California and by Butman et al., (2012) in American rivers not draining peatlands.

4.4. Impact of logging activity on OM fate

- Logging activity can be expected to modify organic matter transformations in peatlands.
- The induced change in hydrology, change in litter quality and addition of extra litter
- 467 (forestry residues) can influence organic matter cycling. In the present case of

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deforestation without engineered drainage, an initial study suggested an increase in peat decomposition in some areas of the deforested site based on the observed increase of the elemental N, S content and change in the OC properties of peat (Gandois et al., 2013). A wider range of δ^{15} N values was found in the peat samples at the deforested site (Table 1). The increase of $\delta^{15}N$ in relation to the increase of N content confirmed the hypothesis that the increase of N content was related to an increase microbial decomposition of the peat at the deforested site. Higher lignin yields ($\lambda 8$) were measured in both peat (+75%) and DOM (+40%) samples in the deforested site compared to the pristine site (Table 2). These higher lignin yields suggested an additional source of lignin at the deforested site: this could reflect the addition of the logging residues to the peat surface (the highest lignin yield has been measured in the Shorea wood samples). In the solid peat samples, no significant difference appeared for the lignin diagenesis indicators between the two sites. If a significant amount of lignin has been added at the deforested site, it has not been transformed to a greater extent than at the pristine site. In contrast, significant differences were observed for both bulk (Δ^{14} C, fluorescence index) and lignin fractions (3,5Bd:V, (Ad/Al)V) of DOM between pristine and deforested sites. A significantly higher DOC concentration at the deforested site was observed (Gandois et al., 2013). The higher Δ^{14} C content of DOM at the deforested site clearly indicated that logging activity has introduced a new DOM source in the peatland porewater. The lower FI indicated less microbial transformation of DOM (McKnight et al. 2001), in agreement with the increased lignin yield. The lignin decomposition indicators were all consistent with a less transformed lignin fraction of DOM at the deforested site, suggesting an input of extra litter that has not reached the characteristic decomposition status of the pristine site.

5. Conclusions

Based on a combination of analytical tools (stable and radiogenic isotopes, lignin biomarkers, optical properties) on both solid and dissolved organic matter in tropical peatlands, this study demonstrates that composition of DOM differs from that of solid organic matter, as indicated by $\delta^{15}N$, $\Delta^{14}C$ and lignin biomarkers. DOM originates to a great extent from living vegetation or litter. Dissolved organic matter is quickly transformed near the peat surface in the upper layer of peat and then advected to deep layers, becoming a source of young carbon deep in the peat. Our study has focused on lignin composition and has demonstrated that lignin is rapidly transformed in the upper layer porewater. Logging activity induced an extra input of lignin in both solid and dissolved phase in the deforested site. DOM composition reflects the change of litter input and organic matter evolution in the deforested site. The analysis of DOM composition is a sensitive tool to assess the impact of change of land use on organic matter cycle, since dissolved organic matter in tropical peatlands is more quickly modified in quantity and composition by changes of land use than is solid organic matter

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7. References

- Anderson J.A.R. (1983) The tropical peat swamps of western Malesia. *Ecosystems of the*
- 520 world 4b: 181-199.

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- Aravena R., B. G. Warner D. J. Charman L. R. Belyea, S. P. Mathur, et H. Dinel. (1993)
- 522 Carbon Isotopic Composition of Deep Carbon Gases in an Ombrogenous Peatland,
- Northwestern Ontario, Canada. Radiocarbon; 35:2.
- Asada T., Warner B. G., Aravena R. (2005) Nitrogen isotope signature variability in plant
- species from open peatland. *Aquatic Botany* **82** (4): 297-307.
- Benner R. Fogel M.L., Sprague E.K., Hodson R.E. (1987) Depletion of 13C in lignin and
- its implications for stable carbon isotope studies. *Nature* **329**: 708-710.
- Benner, Ronald, et Karl Kaiser. 2011. « Biological and Photochemical Transformations
- 529 of Amino Acids and Lignin Phenols in Riverine Dissolved Organic Matter ».
- 530 Biogeochemistry 102 (1-3): 209-222. doi:10.1007/s10533-010-9435-4.

- Billett M. F., Charman D. J., Clark J. M., Evans C. D., Evans M. G., Ostle N. J., Worrall
- F., et al. (2010) Carbon balance of UK peatlands: current state of knowledge and future
- research challenges. *Climate Research* **45**: 13-29.
- Billett M. F., Dinsmore K., J., Smart R.P., Garnett M.H., Holden J., Chapman P., Baird
- 535 A.J., Grayson R., Stott A.W. (2012) Variable Source and Age of Different Forms of
- 536 Carbon Released from Natural Peatland Pipes. Journal of Geophysical Research:
- 537 Biogeosciences 117 (G2): n/a–n/a.
- Boström B., Comstedt D., Ekblad A. (2007) Isotope Fractionation and 13C Enrichment in
- 539 Soil Profiles During the Decomposition of Soil Organic Matter. *Oecologia* 153 (1):
- 540 89-98.
- 541 Butman D., Raymond P.A, Butler K., Aiken G. (2012) Relationships Between Δ14C and
- 542 the Molecular Quality of Dissolved Organic Carbon in Rivers Draining to the Coast from
- 543 the Conterminous United States. *Global Biogeochemical Cycles* **26** (4): n/a–n/a.
- 544 Calvert G.D., Durig J.R., Esterle J.S. (1991) Controls on the chemical variability of peat
- 545 types in a domed peat deposit, Baram River area, Sarawak, Malaysia. International
- 546 *Journal of Coal Geology* **17** (2): 171-188.
- 547 Chanton J. P., Glaser P.H., Chasar L.S., Burdige D.J., Hines M.E., Siegel D.I., Tremblay
- 548 L.B., Cooper W.T. (2008) Radiocarbon Evidence for the Importance of Surface
- 549 Vegetation on Fermentation and Methanogenesis in Contrasting Types of Boreal
- Peatlands. *Global Biogeochemical Cycles* **22** (4): n/a–n/a.
- Chasar L. S., Chanton J. P., Glaser P. H., Siegel D. I., Rivers J. S. (2000) Radiocarbon
- and Stable Carbon Isotopic Evidence for Transport and Transformation of Dissolved

- 553 Organic Carbon, Dissolved Inorganic Carbon, and CH4 in a Northern Minnesota Peatland
- 554 . *Global Biogeochemical Cycles* **14** (4): 1095–1108.
- Cole J. J., Prairie Y. T., Caraco N.F., McDowell W.H., Tranvik L.J., Striegl R.G., Duarte
- 556 C.M., et al. (2007) Plumbing the Global Carbon Cycle: Integrating Inland Waters into the
- Terrestrial Carbon Budget. *Ecosystems* **10** : 172-185.
- 558 Couwenberg J., Dommain R., Joosten H. (2009) Greenhouse gas fluxes from tropical
- peatlands in south-east Asia. *Global Change Biology* (décembre): no-no.
- DeNiro M.J., Hastorf C.A. (1985) Alteration of ¹⁵N14N and 13C12C ratios of plant
- matter during the initial stages of diagenesis: Studies utilizing archaeological specimens
- from Peru. Geochimica et Cosmochimica Acta 49 (1): 97-115.
- Dickens A.F., Gudeman J.A., Gélinas Y., Baldock J.A., Tinner W., Hu F.S., Hedges J.I.
- 564 (2007) Sources and distribution of CuO-derived benzene carboxylic acids in soils and
- 565 sediments. *Organic Geochemistry* **38** (8): 1256-1276.
- 566 Dittmar T., Kattner G. (2003) The biogeochemistry of the river and shelf ecosystem of
- the Arctic Ocean: a review. *Marine Chemistry* **83** (3-4): 103-120.
- Durig D. T., Esterle J. S., Dickson, T. J. Durig J. R. (1988) An Investigation of the
- 569 Chemical Variability of Woody Peat by FT-IR Spectroscopy. Applied Spectroscopy 42
- 570 (7): 1239-1244.
- 571 Ertel J. R., Hedges J.I. (1984) The lignin component of humic substances: Distribution
- among soil and sedimentary humic, fulvic, and base-insoluble fractions. Geochimica et
- 573 *Cosmochimica Acta* **48** (10): 2065-2074.

- 574 Evans C. D, Freeman C., Cork L.G., Thomas D.N, Reynolds R., Billett M.F. Garnert
- 575 M.F, Norris D. (2007) Evidence against recent climate-induced destabilisation of soil
- 576 carbon from 14C analysis of riverine dissolved organic matter. Geophysical research
- 577 *letters* **34** (7).
- 578 Fillion N, Probst A., Probst JL. (1998) Natural organic matter contribution to throughfall
- acidity in French forests ». Environment International 24 (5-6): 547-558.
- Fraser C. J. D., Roulet N. T., Moore T.R. (2001) Hydrology and dissolved organic carbon
- biogeochemistry in an ombrotrophic bog. *Hydrological Processes* **15** (16): 3151-3166.
- Freeman C., Evans C. D., Monteith D. T., Reynolds B., Fenner N. (2001) Export of
- organic carbon from peat soils. *Nature* **412** (6849): 785.
- Gandois L., Cobb A. R., Chieng Hei I., Lim L. B. L., Abu Salim K., Harvey C.F. (2013)
- 585 Impact of Deforestation on Solid and Dissolved Organic Matter Characteristics of
- 586 Tropical Peat Forests: Implications for Carbon Release. *Biogeochemistry*: 1-17.
- 587 Garnett M.H., Hardie S.M.L., Murray C. (2012) Radiocarbon analysis of methane emitted
- from the surface of a raised peat bog. Soil Biology and Biochemistry **50**: 158-163.
- Gleixner G., Danier H. J., Werner R. A., Schmidt H. L. (1993) Correlations Between the
- 590 13C Content of Primary and Secondary Plant Products in Different Cell Compartments
- and That in Decomposing Basidiomycetes. *Plant Physiology* **102** (4): 1287-1290.
- 592 Goñi M.A., Nelson B., Blanchette R.A., Hedges J.I (1993) Fungal degradation of wood
- 593 lignins: Geochemical perspectives from CuO-derived phenolic dimers and monomers.
- *Geochimica et Cosmochimica Acta* **57** (16): 3985-4002.

- 595 Graven H.D., Guilderson T.P., Keeling R.F. (2012) Observations of Radiocarbon in CO2
- 596 at Seven Global Sampling Sites in the Scripps Flask Network: Analysis of Spatial
- 597 Gradients and Seasonal Cycles. Journal of Geophysical Research: Atmospheres 117
- 598 (D2): n/a-n/a.
- 599 Hedges J. I., Mann D.C. (1979) The characterization of plant tissues by their lignin
- oxidation products ». Geochimica et Cosmochimica Acta 43 (11): 1803-1807.
- Hergoualc'h, K., Verchot L.V. (2011) Stocks and fluxes of carbon associated with land
- 602 use change in Southeast Asian tropical peatlands: A review . Global Biogeochemical
- 603 *Cycles* **25**: 13 PP.
- Herne P.J., Bergamaschi B.A., Eckard R.S., Spencer R.G.M. (2009). Fluorescence-based
- proxies for lignin in freshwater dissolved organic matter. Journal of geophysical research
- 606 **114**: G00F3.
- Hernes P. J., Robinson A.C., Aufdenkampe A.K. (2007) Fractionation of Lignin During
- 608 Leaching and Sorption and Implications for Organic Matter "freshness". Geophysical
- 609 Research Letters **34** (17): L17401.
- Hirano T, Jauhiainen J., Inoue T., Takahashi H. (2009) Controls on the Carbon Balance
- of Tropical Peatlands. *Ecosystems* **12** (6) (septembre): 873-887.
- Hou H., Takamatsu T., Koshikawa M. K., Hosomi M. (2005) Copper Complexing
- 613 Capacity of Throughfall and its Environmental Effect. Water, Air, & Soil Pollution 162
- 614 (1): 229-245.
- Kalbitz, K. (2001) Properties of organic matter in soil solution in a German fen area as
- dependent on land use and depth. *Geoderma* **104** (3-4): 203-214.

- Kobayashi, S. (1999) Initial phase of secondary succession in the exploited peat swamp
- 618 forest (Shorea albida) at sungai Damit, Belait in Brunei Darussalam. Proceedings of the
- 619 international symposium on tropical peatlands. Bogor, Indonesia, 22-23 November 1999.
- 620 205-214.
- 621 Kokfelt, U., P. Rosén, K. Schoning, T. R. Christensen, J. Förster, J. Karlsson, N. Reuss,
- et al. 2009. « Ecosystem Responses to Increased Precipitation and Permafrost Decay in
- 623 Subarctic Sweden Inferred from Peat and Lake Sediments ». Global Change Biology 15
- 624 (7): 1652-63. doi:10.1111/j.1365-2486.2009.01880.x.
- Kramer M. G., Sollins P., Sletten R.S, Swart P.K. (2003) N Isotope fractionation and
- measures of organic matter alteration during decomposition. *Ecology* **84** (8): 2021-2025.
- 627 Limpens J., Berendse F., Blodau C., Canadell J. G., Freeman C., Holden J., Roulet N.,
- 628 Rydin H., Schaepman-Strub G. (2008) Peatlands and the carbon cycle: from local
- 629 processes to global implications a synthesis. *Biogeosciences* **5** (5): 1475-1491.
- Malmer, Nils, et Elis Holm, 1984, « Variation in the C/N-Ouotient of Peat in Relation to
- Decomposition Rate and Age Determination with 210 Pb ». Oikos 43 (2): 171-182.
- 632 doi:10.2307/3544766.
- Martínez C.A., Biester H., Mighall T., Bindler R. (2007) Climate-driven enrichment of
- 634 pollutants in peatlands. *Biogeosciences* **4**: 905-911.
- McKnight D.M., Boyer E.W., Westerhoff P.K., Doran P.T. Kulbe T., Andersen D.T.
- 636 (2001) Spectrofluorometric characterization of dissolved organic matter for indication of
- precursor organic material and aromaticity. *Limnology and oceanography* **46** (1): 38-48.

- 638 Miettinen J., Shi C., Liew S.C. (2011) Deforestation Rates in Insular Southeast Asia
- 639 Between 2000 and 2010. *Global Change Biology* **17** (7): 2261-2270.
- Moore S., Gauci V., Evans C.D, Page S.E. (2011) Fluvial organic carbon losses from a
- Bornean blackwater river. *Biogeosciences* **8**: 901-909.
- Natelhoffer K. J., Fry B. (1988) Controls on Natural Nitrogen-15 and Carbon-13
- 643 Abundances in Forest Soil Organic Matter. Soil Science Society of America Journal 52
- 644 (6): 1633-1640.
- Neff J.C., Asner G.P. (2001) Dissolved Organic Carbon in Terrestrial Ecosystems:
- 646 Synthesis and a Model. *Ecosystems* **4** (1): 29-48.
- Page S.E., Rieley J.O, Banks C.J. (2011) Global and regional importance of the tropical
- peatland carbon pool. *Global Change Biology* **17**: 798-818.
- Pastor J., Solin J., Bridgham S.D, Updegraff K., Harth C., Weishampel P., Dewey B
- 650 (2003) Global Warming and the Export of Dissolved Organic Carbon from Boreal
- 651 Peatlands. *Oikos* **100** (2): 380-386.
- Posa M. R. C., Wijedasa L.S, Corlett R.T. (2011) Biodiversity and Conservation of
- Tropical Peat Swamp Forests. *BioScience* **61** (1): 49-57.
- Powers J.S., Schlesinger W.H. (2002) Geographic and vertical patterns of stable carbon
- isotopes in tropical rain forest soils of Costa Rica. *Geoderma* **109** (1–2): 141-160.
- Raymond P. A., McClelland J.W., Holmes R.M., Zhulidov A.V, Mull K., Peterson B.J.,
- 657 Striegl R.G, Aiken G.R, Gurtovaya T.Y. (2007) Flux and Age of Dissolved Organic
- 658 Carbon Exported to the Arctic Ocean: A Carbon Isotopic Study of the Five Largest Arctic
- Rivers. Global Biogeochemical Cycles 21 (4): n/a–n/a.

- Sanderman J., Lohse K.A., Baldock J.A, AmundsonR. (2009) Linking Soils and Streams:
- Sources and Chemistry of Dissolved Organic Matter in a Small Coastal Watershed. Water
- 662 *Resources Research* **45** (3): n/a–n/a.
- ŠantRůČková H., Bird M. I., Lloyd J. (2000) Microbial Processes and Carbon-isotope
- 664 Fractionation in Tropical and Temperate Grassland Soils. Functional Ecology 14 (1):
- 665 108–114.
- Spencer R.G.M., Hernes P.J., Aufdenkampe A.K., Baker A., Gulliver P., Stubbins A.,
- Aiken G.R, (2012). An initial investigation into the organic matter biogeochemistry of
- the Congo River. *Geochimica et Cosmochimica Acta* **84**: 614-627.
- Tareq S. M., Tanaka N., Ohta K. (2004) Biomarker signature in tropical wetland: lignin
- 670 phenol vegetation index (LPVI) and its implications for reconstructing the
- paleoenvironment Science of The Total Environment **324** (1-3): 91-103.
- 672 Tfaily M. M., Hamdan R., Corbett J.E, Chanton J.P, Glaser P.H., Cooper W.T. (2013)
- 673 Investigating dissolved organic matter decomposition in northern peatlands using
- 674 complimentary analytical techniques. *Geochimica et Cosmochimica Acta* 112: 116-129.
- Thevenot, Mathieu, Marie-France Dignac, et Cornelia Rumpel. 2010. « Fate of lignins in
- 676 soils: A review ». Soil Biology and Biochemistry 42 (8): 1200-1211.
- 677 doi:10.1016/j.soilbio.2010.03.017.
- Weishaar J.L, Aiken G.R, Bergamaschi B.A., Fram M.S., Fujii R., Mopper K. (2003)
- 679 Evaluation of specific ultraviolet absorbance as an indicator of the chemical composition
- and reactivity of dissolved organic carbon. *Environmental Science & Technology* **37** (20):
- 681 4702-4708.

- Williams, C.J. Yavitt J.B. (2003) Botanical composition of peat and degree of peat
- decomposition in three temperate peatlands ». *Ecoscience* **10**(1):85-95.
- Williams C.J, Yavitt J.B, Wieder K., Cleavitt N.L. (1998) Cupric oxide oxidation
- products of northern peat and peat-forming plants. Canadian Journal of Botany 76 (1):
- 686 51-62.
- 687 Yamashita Y., Scinto L.J., Maie N., Jaffé R. (2010) Dissolved Organic Matter
- 688 Characteristics Across a Subtropical Wetland's Landscape: Application of Optical
- Properties in the Assessment of Environmental Dynamics. *Ecosystems*.
- 690 Yule C., Gomez L. (2009) Leaf litter decomposition in a tropical peat swamp forest in
- 691 Peninsular Malaysia. Wetlands Ecology and Management 17 (3): 231-241.

8. Captions

- Table 1. Bulk analysis of solid and dissolved organic matter. Mean (min/max) values
- Table 2. CuO oxidation products. V, organic carbon normalized yields of the vanillyl
- 695 phenols (VI, vanillin; Vn, acetovanillone; Vd, vanillic acid); S, organic carbon
- normalized yields of the syringyl phenols (Sl, syringealdehyde; Sn, acetosyringone; Sd,
- 697 syringic acid); C, organic carbon normalized yields of the cinnamyl phenols (p-Cd, p-
- 698 coumaric acid; Fd: ferulic acid); P, carbon normalized yields of p-hydroxyphenols (Pl, p-
- 699 hydroxybenzaldehyde; Pn, p-hydroxyacetophenone; Pd, p-hydroxybenzoic acid); 3,5Bd,
- 700 3,5-dihydroxybenzoic acid. All the content are expressed in mg 100 mg OC⁻¹. Mean
- 701 (min/max) values are indicated.

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- 703 Figure 1: Map of Borneo Island, Brunei Darussalam and the study area (Ulu Mendaram
- 704 Conservation Area). Study sites are two domes, located on both sides of the Mendaram
- 705 river.
- Figure 2: (a) δ^{13} C (‰) content of vegetation, peat and DOM samples of porewater and
- 707 river water. (b): $\delta^{15}N$ (%) in relation to C/N ratio of solid and dissolved organic matter.
- Figure 3: δ^{14} C (‰) content versus depth for peat cores and DOM samples.
- 709 Figure 4: S:V vs C:V content of vegetation, peat and DOM samples of porewater and
- 710 river water.
- Figure 5: P:(V+S) vs 3,5Bd:V content of vegetation, peat and DOM samples of porewater
- and river water.
- 713 Figure 6: (Ad/Al)V vs (Ad/Al)S content of vegetation, peat and DOM samples of
- 714 porewater and river water.
- 715 Figure 7: C:V (a) and Δ^{14} C (b) relationship with fluorescence index.

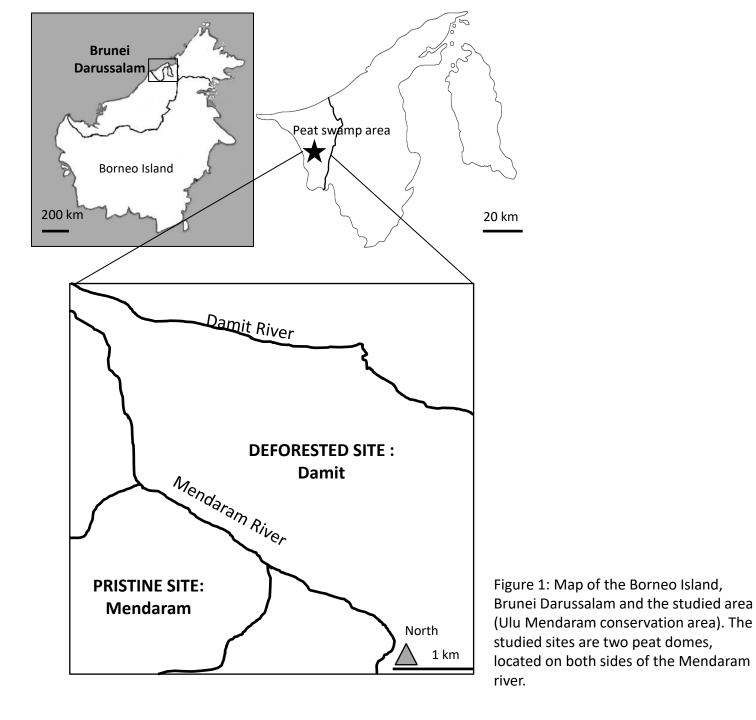
716 Table 1. Bulk analysis of solid and dissolved organic matter. Mean (min/max) values.

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Nom	n	С	N	$\delta^{13}C$	$\delta^{15}N$	Δ^{14} C	DOC	SUVA ₂₈₀	SUVA350	FI
		(%)	(%)	(‰)	(‰)	(‰)	(mg L ⁻¹)	(L m ⁻¹ ·mg C ⁻¹)		-
DOM_River_Pristine	2	42.4(41.7/43.1)	1.08(1.06/1.11)	-29.66(-29.7/-29.62)	-2.37(-3.35/-1.39)	68.28 (62.73-73.84)	37.4 (37.3-37.4)	4.48(4.42/4.53)		1.50(1.50/1.49)
DOM_River_Deforested	1	31.3	0.84	-29.69	-7.5		26.7	4.00		1.62
DOM _ Pristine	5	41.8(40.7/43.14)	0.90(0.67/1.04)	-29.9(-30.08/-29.47)	-4.18(-7.4/-0.79)	74.58 (70.87-77.22)	66.1 (59.7 -81.4)	4.28(3.43/4.63)		1.41(1.39/1.46)
DOM_deforested	7	44.5(40.2/47.27)	0,10(0.6/1.19)	30.18(-30.50/-29.84)	-4.13(-8.3/-1.3)	87.84 (83.77-94.4)	74.9 (73.5-79.4)	5,16(4.38/5.99)		1.36(1.33/1.43)
Shorea	3	50.6(49.5/52.7)	0.98(0.4/1.38)	-29.69(-29.82/-29.59)	-6(-7.88/-3.1)					
Pandan	2	47.35(47.3/47.52)	1.12(0.7/1.2)	-29.67(-29.69/-29.66)	-1.82(-7.46/-1.46)					
Fern	2	43.34(43.1/43.53)	2.67(2.6/3.18)	-30.87(-31/-30.75)	-5.31(-5.31/-4.91)					
Peat_pristine	9	52.4(49.6/55.5)	1.95(1.68/2.18)	30.49(-32.32/-29.79)	0.92(-1.72/-0.22)					
Peat_deforested	9	50.95(48.1/55.5)	2.09(1.33/2.72)	29.48(-30.65/-28.32)	1.22(-2.74/0.91)					

Table 2 . CuO oxidation products. Mean (min/max) values. ΣS = syringealdehyde (Sl)+ acetosyrigon (Sn) + syringic acid (Sd); ΣV= vanillin (Vl) + acetovanillon (Vn) + vanillic acid (Vd); ΣC= coumaric acid, (p-Cd) + ferulic acid (Fd); ΣP= p-hydroxybenzaldehyde (Pl)+ p-hydroxyacetophenon (Pn)+p-hydroxybenzoic acid (Pd). 3,5Bd= 3,5-dihydroxybenzoic acid (3,5Bd). (Ad/Al)S = syringic acid/syringealdehyde, (Ad/Al)V= vanillic acid/vanillin.

Nom	n	λ8	ΣS	ΣV	ΣC	ΣΡ	3-5Bd:V	(Ad/Al)S	(Ad/Al)V
		mg 100mgOC ⁻¹	mg.100mgOC ⁻¹	mg.100mgOC ⁻¹	mg.100mgOC ⁻¹	mg.100mgOC ⁻¹			
DOM_River_Pristine	2	1.25(0.96/1.53)	0.6(0.4/0.8)	0.5(0.4/0.6)	0.031(0.02/0.04)	0.9(0.69/1.13)	1.343(1.24/1.44)	0.91(0.70/1.12)	1.6(1.57/1.64)
DOM_River_Deforested	1	1.63	0.9	0.6	0.04	1.28	1.26	1.48	0.95
DOM _ Pristine	5	1.16(0.80/1.50)	0.6(0.3/0.8)	0.5(0.3/0.6)	0.019(0.01/0.05)	1.14(1.05/1.3)	1.58(1.31/1.98)	1.16(0.82/1.5)	1.42(1.13/1.65)
DOM_Deforested	7	1.51(0.65/1.99)	0.7(0.2/0.8)	0.8(0.3/1.2)	0.03(0.01/0.03)	1.22(0.74/1.80)	0,92(0.75/1.17)	1.18(1.07/1.27)	1.05(0.86/1.17)
Shorea	3	10.70(8.18/12.48)	4.6(4.1/5)	5.6(3/7.7)	0.44(0.03/1.0)	0.96(0.14/1.85)	0.049(0.01/0.11)	0.39(0.38/0.41)	0.45(0.36/0.62)
Pandanus	2	9.15(8.11/10.18)	5.1(4.6/5.1)	3.9(3.9/5)	0.067(0.05/0.08)	0.29(0.27/0.31)	0.66(0.58/0.74)	0.41(0.41/0.42)	0.39(0.36/0.42)
Fern	2	3.55(0.96/6.14)	0.1(0.1/0.1)	3.3(0.6/6)	0.102(0.1/0.1)	0.80(0.8/1.1)	0.09(0.03/0.15)	1.1(1.05/1.15)	0.42(0.42/0.45)
Peat_pristine	9	5.40(4.04/10.49)	2.2(1.1/5.7)	3(2.1/4.5)	0.17(0.01/0.31)	1.55(0.84/3.56)	0.12(0.07/0.19)	0.57(0.43/0.69)	0.74(0.54/1)
Peat_deforested	9	9.55(4.60/18.97)	4.2(1.4/9.6)	5.1(2.5/9)	0.23(0.15/0.54)	2.18(0.77/4.05)	0.08(0.03/0.16)	0.55(0.43/0.72)	0.66(0.52/0.85)



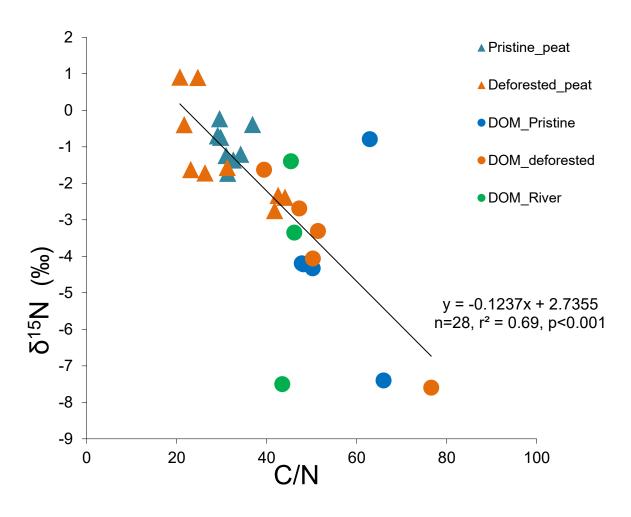


Figure 2: $\delta^{15}N$ (‰) in relation to C/N ratio of solid (peat) and dissolved organic matter (DOM).

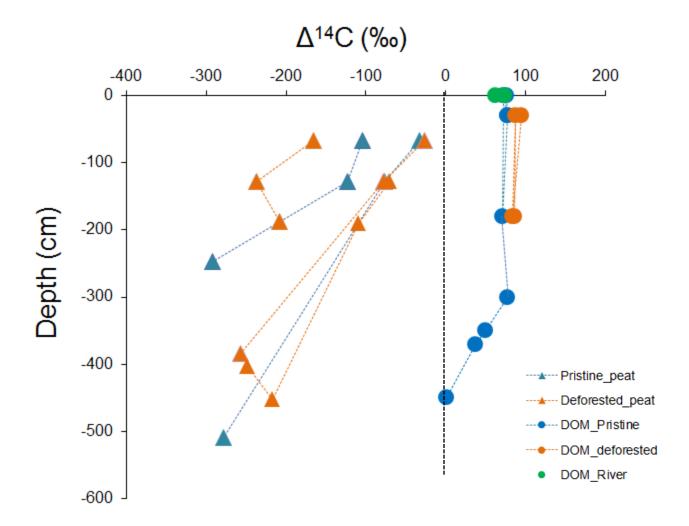


Figure 3: Δ^{14} C (‰) content versus depth for peat cores and DOM samples. Positive values indicate Modern DOM.

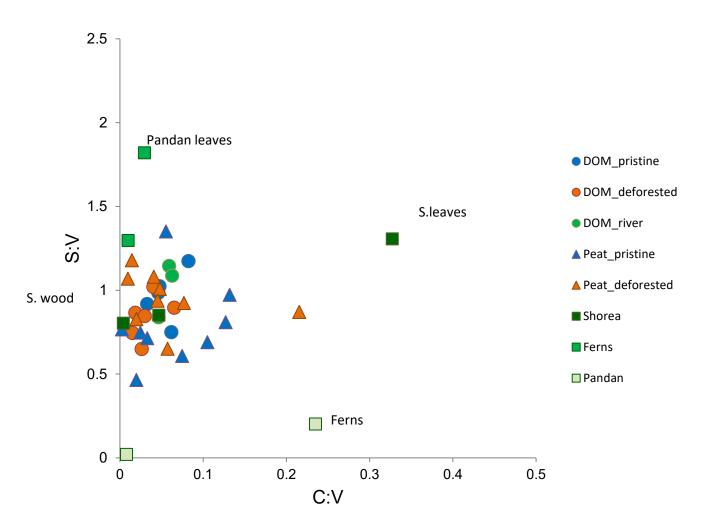
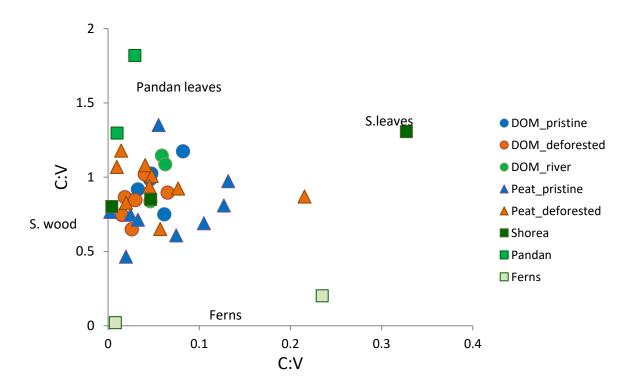


Figure 4a: S/V vs C/V content of vegetation, peat and DOM samples of pore water and river water.



OR: Figure 4a: S/V vs C/V content of vegetation, peat and DOM samples of pore water and river water.

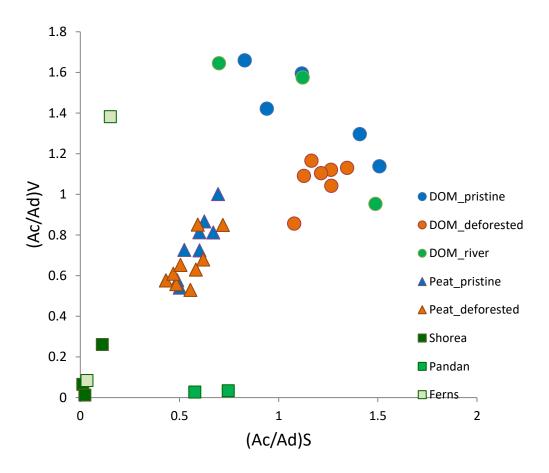


Figure 4c: (Ac/Ad)V vs (Ac/Ad)S content of vegetation, peat and DOM samples of pore water and river water.

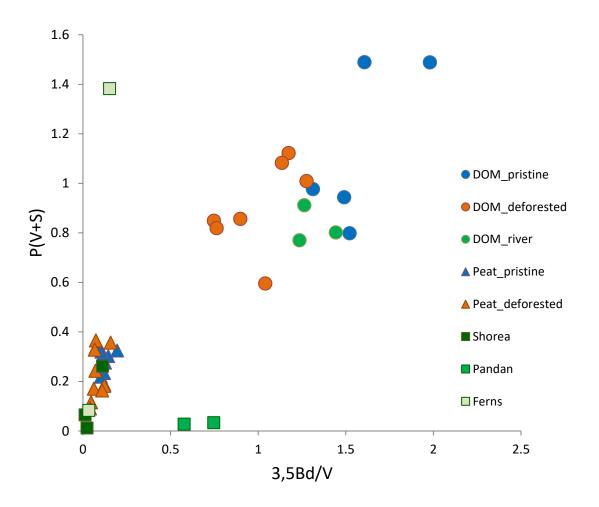


Figure 4b: P/(V+S) vs 3,5Bd/V content of vegetation, peat and DOM samples of pore water and river water.

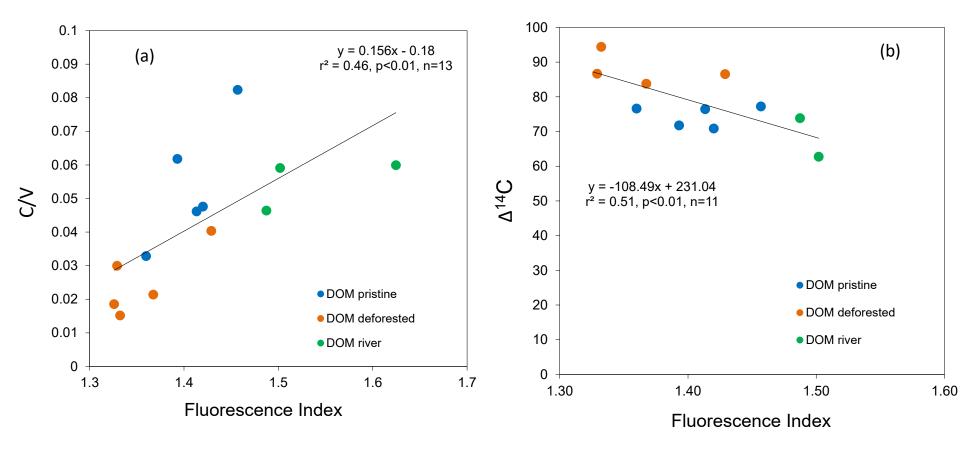


Figure 5: C/V (a) and Δ^{14} C (b) relationship with fluorescence index.